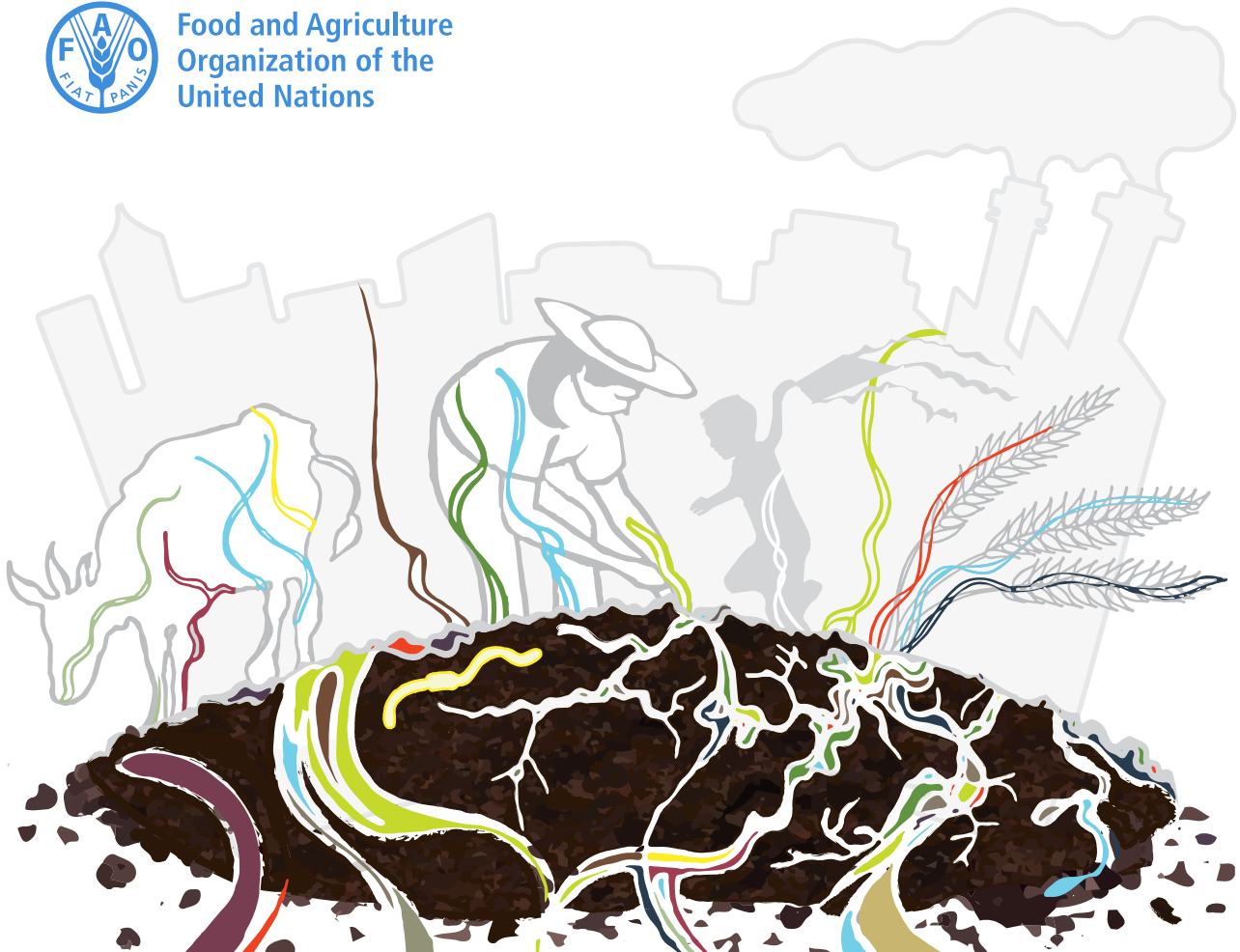




Food and Agriculture  
Organization of the  
United Nations



# SOIL POLLUTION: A HIDDEN REALITY







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## EXECUTIVE SUMMARY

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“Soil pollution” refers to the presence in the soil of a chemical or substance out of place and/or present at a **higher than normal concentration** that has adverse effects on any non-targeted organism. Soil pollution often cannot be directly assessed or visually perceived, making it a hidden danger.

The Status of the World's Soil Resources Report (SWSR) identified soil pollution as one of the main soil threats affecting global soils and the ecosystems services provided by them.

Concerns about soil pollution are growing in every region. Recently, the United Nations Environmental Assembly (UNEA-3) adopted a resolution calling for accelerated actions and collaboration to address and manage soil pollution. This consensus, achieved by more than 170 countries, is a clear sign of the global relevance of soil pollution and of the willingness of these countries to develop concrete solutions to address the causes and impacts of this major threat.

The main anthropogenic sources of soil pollution are the chemicals used in or produced as byproducts of industrial activities, domestic, livestock and municipal wastes (including wastewater), agrochemicals, and petroleum-derived products. These chemicals are released to the environment accidentally, for example from oil spills or leaching from landfills, or intentionally, as is the case with the use of fertilizers and pesticides, irrigation with untreated wastewater, or land application of sewage sludge. Soil pollution also results from atmospheric deposition from smelting, transportation, spray drift from pesticide applications and incomplete combustion of many substances as well as radionuclide deposition from atmospheric weapons testing and nuclear accidents. New concerns are being raised about emerging pollutants such as pharmaceuticals, endocrine disruptors, hormones and toxins, among others, and biological pollutants, such as micropollutants in soils, which include bacteria and viruses.

Based on scientific evidence, soil pollution can severely degrade the major ecosystem services provided by soil. Soil pollution reduces food security by both reducing crop yields due to toxic levels of contaminants and by causing crops produced from polluted soils to be unsafe for consumption by animals and humans. Many contaminants (including major nutrients such as nitrogen and phosphorus) are transported from the soil to surface waters and ground water, causing great environmental harm through eutrophication and direct human health issues due to polluted drinking water. Pollutants also directly harm soil microorganisms and larger soil-dwelling organisms and hence affect soil biodiversity and the services provided by the affected organisms.

The results of scientific research demonstrate that soil pollution directly affects human health. Risks to human health arise from contamination from elements such as arsenic, lead, and cadmium, organic chemicals such as PCBs (polychlorinated biphenyls) and PAHs (polycyclic aromatic hydrocarbons), and pharmaceuticals such as antibiotics. The health risks associated with the widespread soil contamination by radionuclides from the Chernobyl disaster in 1986 are an enduring memory for many people.



Remediation of polluted soils is essential, and research continues to develop novel, science-based remediation methods. Risk assessment approaches are similar worldwide and consist of a series of steps to be taken to identify and evaluate whether natural or human-made substances are responsible for polluting the soil, and the extent to which that pollution is posing a risk to the environment and to human health. Increasingly expensive physical remediation methods such as chemical inactivation or sequestration in landfills are being replaced by science-based biological methods such as enhanced microbial degradation or phytoremediation.

FAO's *Revised World Soil Charter* recommends that national governments implement regulations on soil pollution and limit the accumulation of contaminants beyond established levels in order to guarantee human health and wellbeing, a healthy environment and safe food. Governments are also urged to facilitate remediation of contaminated soils that exceed levels established to protect the health of humans and the environment. It is also essential to limit pollution from agricultural sources by the global implementation of sustainable soil management practices.

This book aims to summarise the state of the art of soil pollution, and to identify the main pollutants and their sources affecting human health and the environment, paying special attention to those pollutants that are present in agricultural systems and that reach humans through the food chain. It concludes with some case studies of the best available techniques for assessing and remediating contaminated soils.

This book has been developed within the framework of the Global Symposium on Soil Pollution (GSOP18), identifying the main gaps in knowledge on soil pollution worldwide and serving as a basis for future discussions.



## GLOSSARY

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**Contaminant:** substance or agent present in the soil as a result of human activity (ISO, 2013).

**Leaching:** the dissolution and movement of dissolved substances by water (ISO, 2013).

**Parent material:** The original material (mineral and/or organic) from which soil developed by pedogenetic processes.

**Persistent organic pollutant (POP):** Synthesized carbon-based compounds from agrochemicals and industrial products that generally biodegrade very poorly and most of which will bioaccumulate in tissues of organisms. Some pesticides are POPs, as are Polychlorinated dibenzodioxins (PCDDs), Polychlorinated dibenzofurans (PCDFs), Polychlorinated biphenyls (PCBs), and Polycyclic aromatic hydrocarbons (PAHs).

**Soil:** the upper layer of the Earth's crust transformed by weathering and physical/chemical and biological processes. It is composed of mineral particles, organic matter, water, air and living organisms organized in genetic soil horizons (ISO, 2013).

**Soil ecosystem functions:** description of the significance of soils to humans and the environment. Examples are: (1) control of substance and energy cycles within ecosystems; (2) basis for the life of plants, animals and man; (3) basis for the stability of buildings and roads; (4) basis for agriculture and forestry; (5) carrier of genetic reservoir; (6) document of natural history; and (7) archaeological and paleo-ecological document (ISO, 2013).

**Soil health:** the continued capacity of the soil to function as a vital living system, within ecosystem and land-use boundaries, to sustain biological productivity, promote the quality of air and water environments, and maintain plant, animal, and human health (Doran, Stamatiadis and Haberern, 2002).

**Soil ecosystem services:** the capacity of natural processes and components to provide goods and services that satisfy human needs, directly or indirectly (Groot, 1992).

**Food security:** it is defined as the availability, access, utilization and stability of food supply.

**Soil contamination:** occurs when the concentration of a chemical or substance is higher than would occur naturally but is not necessarily causing harm (this volume).

**Soil pollution:** refers to the presence of a chemical or substance out of place and/or present at higher than normal concentration that has adverse effects on any non-targeted organism (this volume).



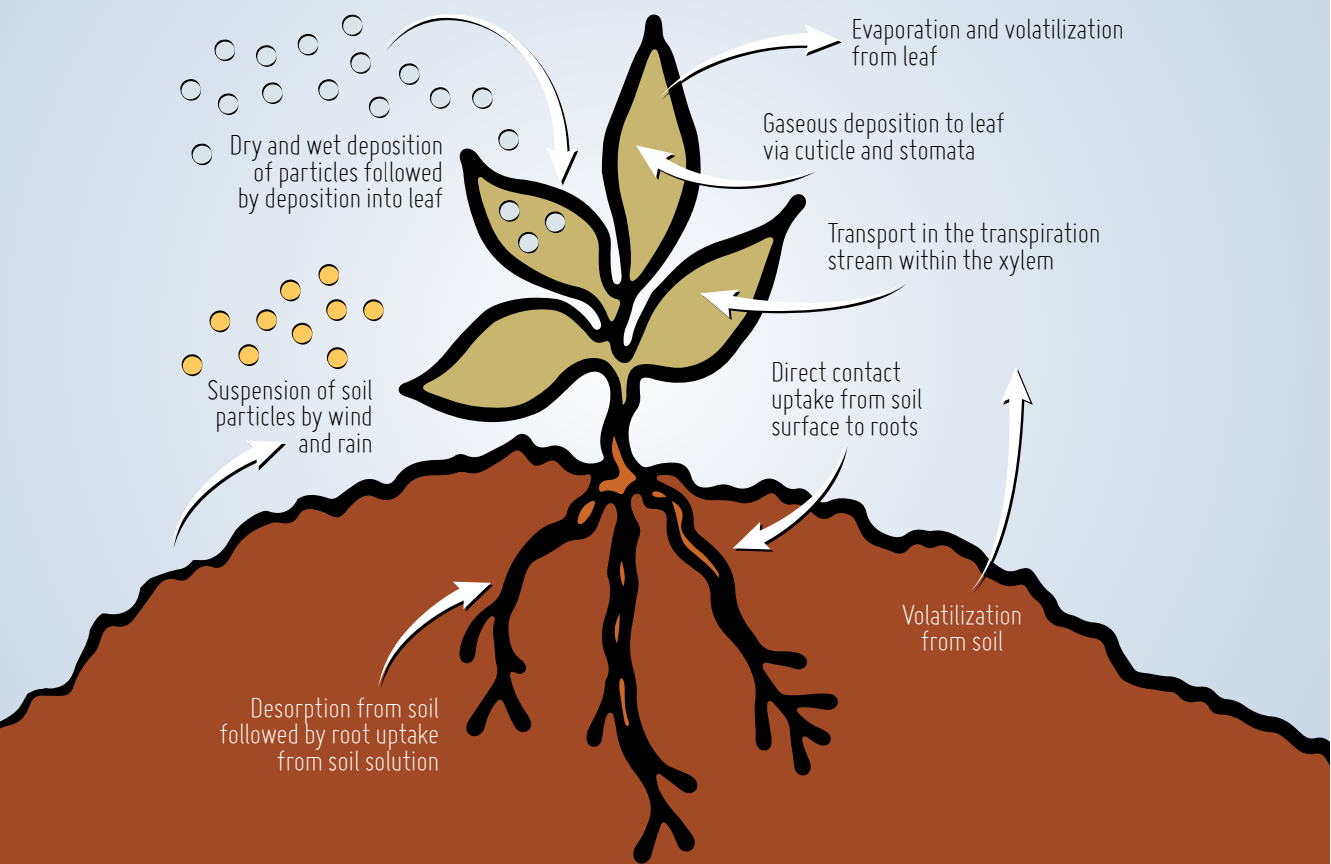


## 2 | THE IMPACTS OF SOIL POLLUTION ON THE FOOD CHAIN AND ECOSYSTEM SERVICES

The predicted world's population of over nine billion by 2050 will require the provision of enough good quality food and water (Godfray *et al.*, 2010; McBratney, Field and Koch, 2014). According to Dubois (Dubois, 2011), food production will increase by 70 percent by 2050 globally, and by 100 percent in developing countries, compared with 2009 production levels. FAO's latest projections indicate that global food production will increase by 60 percent between 2005/07 and 2050 under its baseline scenario. This represents a downward revision, based on updated data and information, from the 70 percent increase projected for the same period in 2009. (World Agriculture Towards 2030/2050: The 2012 revision ESA E Working PaperNo. 12-03 <http://www.fao.org/economic/esa/esag/en/>). The quantity and nutritional quality of food supports human health, and 95 percent of food production depends on soils (Oliver and Gregory, 2015; FAO, 2015). Only healthy soils can provide the needed ecosystem services and secure supplies of more food and fibre. The provision of ecosystem services has received considerable attention and can be defined as "the capacity of natural processes and components to provide goods and services that satisfy human needs, directly or indirectly" (Groot, 1992). Food security is defined as "the availability, access, utilization and stability of food supply." Soil pollution reduces food security both by reducing crop yields due to toxic levels of contaminants and by causing the produced crops to be unsafe for consumption (FAO and ITPS, 2015).

## 2.1 | SOIL POLLUTION, PLANT UPTAKE AND FOOD CHAIN CONTAMINATION

The pathways of contamination within the food chain by the transfer of soil pollutants through plants are shown in Figure 14.



**Figure 14.** Principal uptake pathways for the uptake of soil contaminants by plants (adapted from Collins, Fryer and Grosso, 2006)

If a contaminant is highly toxic to plants at low concentrations and is not easily translocated to shoots, fruits or tubers to pose a hazard to animals and humans, it is unlikely to enter the food chain and become a hazard. This concept was termed the “Soil-Plant Barrier” by Chaney almost 40 years ago for metals and metalloids (Chaney, 1980). Chaney defined four groups of metals entering the food chain when sewage sludge was applied to soil, as a function of their danger to human health (Table 6).

**Table 6.** Metals/metalloids classified in groups according to potential food-chain risk via plant uptake.  
Adapted from: Chaney, 1980

Group 1	Group 2	Group 3	Group 4
Silver (Ag)	Mercury (Hg)	Boron (B)	Arsenic (As)
Chromium (Cr)	Lead (Pb)	Copper (Cu)	Cadmium (Cd)
Tin (Sn)		Manganese (Mn)	Cobalt (Co)
Titanium (Ti)		Molybdenum (Mo)	Molybdenum (Mo)
Yttrium (Y)		Nickel (Ni)	Selenium (Se)
Zirconium (Zr)		Zinc (Zn)	Thallium (Tl)

Group 1 comprises the elements that pose a low risk of food chain contamination because they are not taken up by plants, due to their low solubility in soil, which means negligible uptake and translocation by plants. Elevated concentrations of these elements in foods usually indicate direct contamination through soil or dust accumulation. Group 2 includes elements that are strongly sorbed by soil surfaces, and while they may be absorbed by plant roots, they are not readily translocated to the edible tissues and therefore pose minimal risks to human health. These elements could, however, pose a risk to grazing animals (or humans) if contaminated soil is ingested. Group 3 comprises the elements that are readily taken up by plants, but that are phytotoxic at concentrations that pose little risk to human health. Conceptually, the “soil-plant barrier” protects the food chain from contamination by these elements. Group 4 consists of elements that are at the highest risk for food-chain contamination as they pose risks to human or animal health at plant tissue concentrations that are not generally phytotoxic. Chaney originally classified As in Group 2, but research over the last 20 years has indicated that flooded rice systems are at risk from As transfer through the food chain due to low redox conditions in flooded soils. This increases the solubility of As for uptake by rice and hence As should now be classified as a high-risk, Group 4 element. Contamination of soils by As and Cd is perhaps the most widespread risk to the food chain globally (Grant *et al.*, 1999; McLaughlin, Parker and Clarke, 1999), with large areas of South-East Asia having soils contaminated by As (Meharg, 2004) and Cd (Hu, Cheng and Tao, 2016).

In some parts of China, soils polluted by heavy metals are nevertheless used to grow grain. The grain grown in these soils is often in turn polluted with heavy metals. According to China Dialogue, an estimated 12 million tonnes of polluted grain must be disposed of each year, costing Chinese farmers up to CNY 20 billion, or about USD 2.57 billion (Lynn, 2017).

Excess heavy metals such as arsenic, cadmium, lead and mercury in soils can also impair plant metabolism and decrease crop productivity, ultimately putting pressure on arable land. When they enter the food chain, these pollutants also

pose risks to food security, water resources, rural livelihoods and human health. The uptake and translocation of metals into above ground tissues are conditioned by genetic and physiological differences of plants (Chen, Li and Shen, 2004), as well as by the concentration of metals in the soil and the exposure time (Rizwan *et al.*, 2017; Tózsér, Magura and Simon, 2017). Once metals enter plant tissues, they may interfere with several metabolic processes, reducing plant growth and causing toxicity and finally plant death. Decreases in germination rates, oxidative damage, lower roots and shoots elongation and alterations of sugar and protein metabolisms were the main effects reported (Ahmad and Ashraf, 2011). High levels of lead, for example, accelerate the production of reactive oxygen species, causing lipid membrane and chlorophyll damage that further leads to the alteration of photosynthetic processes and of the overall growth of the plant (Najeeb *et al.*, 2017). Cadmium can accumulate in different edible tissues (Baldantoni *et al.*, 2016), causing reduction of root, stem and leaf growth, decreasing net photosynthesis and water use efficiency and altering nutrient uptake (Rizwan *et al.*, 2017).

Radionuclides may also present a potential threat to food quality, through atmospheric deposition of radionuclides on soil from nuclear energy accidents, or the addition of radionuclides to the soil through fertilizers or through wastes and byproducts from the nuclear industry (Mortvedt, 1994). The potential for transfer of radionuclides from soil to plants and to the food chain was first identified in the 1950s both in restricted areas where nuclear weapons testing had taken place and more generally through deposition of fallout from bomb testing. The Chernobyl nuclear accident in 1986 caused widespread pollution of soils by radionuclides (principally <sup>131</sup>I, <sup>134</sup>Cs and <sup>137</sup>Cs) (Bell, Minski and Grogan, 1988). Uptake of pollutants from soils into forages was followed by the contamination of grazing animals and led to restrictions being placed on the sale and slaughter of sheep from affected areas in the United Kingdom of Great Britain and Northern Ireland (Smith *et al.*, 2000). Widespread contamination of the food chain by the same radionuclides also occurred after the Fukushima nuclear accident in Japan (Berends and Kobayashi, 2012).

Compared to metals, metalloids and radionuclides, the global footprint of soils that are highly contaminated by organic contaminants is much smaller, with contamination of the food chain being localized around industrial or urban centres. Contamination occurs through waste re-use on land and as a legacy of the use or disposal of persistent and bioaccumulative organic chemicals (principally organochlorines (OCs), polychlorinated dibenzo-p-dioxin (PCDD), polychlorinated dibenzofurans (PCDF) and poly- and perfluoroalkyl substances (PFAS)). Levels of soil contamination are usually lower than for metals/metalloids, generally in concentrations less than mg/kg in soil, especially for dioxins, furans and PFAS compounds.

The pathway of uptake of organic contaminants into the food chain depends on the properties of the organic contaminant – principally their volatility, hydrophobicity and solubility in water. Hydrophilic organic contaminants with low volatility (e.g. PFAS) will principally enter the food chain through root uptake and translocation to food parts (Navarro *et al.*, 2017). On the other hand, volatile hydrophobic contaminants (e.g. dioxins, furans and polychlorinated biphenyls (PCBs)) will tend to accumulate in the food chain through atmospheric uptake, as they are generally strongly sorbed in soils (Collins, Fryer and Grosso, 2006; Simonich and Hites, 1995). Some plant species, however, can accumulate these compounds through uptake from soil (Huelster, Mueller and Marschner, 1994). Sun *et al.*, were able to quantify organic



pollutants and their residues in plants by measuring the uptake of <sup>14</sup>C-residues (Sun *et al.*, 2018). Many studies have demonstrated the uptake of pesticide residues through several main pathways. These residues are incorporated in plant tissues and transferred to the end consumer (Randhawa *et al.*, 2014). However, unlike metals and metalloids, the incidence and severity of adverse effects on human health caused by soil contamination by organic chemicals is much less well documented or demonstrated, likely because smaller areas of land are affected by this type of pollutant, and contamination levels are generally lower.

Atmospheric deposition and gas exchange through the stomata are the main mechanisms of absorption of POPs by plants; the pollutants are later translocated to other plant tissues and accumulated in their hydrophobic lipids and waxes (Odabasi *et al.*, 2016). Root uptake of POPs from soils is limited since POPs are tightly bound to soil particles (Collins, Fryer and Grosso, 2006). For that reason, soils can be viewed as a reservoir and as a source of POPs with low and medium volatility; these may enter the food chain after being taken up by plants from the atmosphere. Absorption efficiency of POPs across the gastro-intestinal tract and their storage and release dynamics are intimately linked to that of fat storage and metabolism (Sweetman, Thomas and Jones, 1999).

## 2.2 | IMPACT ON ECOSYSTEM SERVICES OF SOIL POLLUTION FROM AGRICULTURE

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Entry of pollutants directly (release of effluents on land) or indirectly (use of polluted water as irrigation to crops) has been reported to contaminate vast areas of soil resources and groundwater bodies, affecting crop production as well as human and animal health through food contamination (Saha *et al.*, 2017).

Agricultural inputs such as fertilizers, pesticides, antibiotics contained in animal manure or the ones used for illness prevention and infection treatment in plants are major potential pollutants in agricultural lands and pose special challenges due to the fast-changing chemical formulas employed (GSP, 2017). Intensification of agriculture to produce enough food, fibre and biofuel has led to a heritage of polluted soils. In China, heavy metals contents have considerably increased in the last 30 years, with values oscillating between 48 percent for Zn to over 250 percent when Cd is compared with its background levels in 1990 (Zeng, Li and Mei, 2008). However, the transfer of pollutants from soil to plants is not yet well understood, and the question “is food produced in healthier soils also more nutritious?” needs stronger scientific evidence to engage policy makers, governments and land users towards sustainable and environmentally friendly practices and to leave behind more business-oriented approaches.

### 2.2.1 | SYNTHETIC FERTILIZERS

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Modern agriculture practices accelerate soil pollution with the intensive use of fertilizer and pesticides in order to increase productivity and reduce crop losses. When pollutants reach high levels in the soil, not only do soil degradation processes take place, but crop productivity can also be affected. Therefore, in addition to endangering human health and the environment, soil pollution can also cause economic losses.

Excess N in soil has been identified as the main cause of soil acidification and salinization through nitrification and other N-transformation processes. Soils acidify very slowly under natural conditions over hundreds to millions of years (Guo *et al.*, 2010), but this process is significantly accelerated by agricultural practices, mainly excessive N fertilization, which causes reductions in soil pH by 0.26 pH units on average in different land uses (Lucas *et al.*, 2011; Tian *et al.*, 2015; Zhao *et al.*, 2014a). The analysis of acidification sources in agricultural soils in China demonstrated that anthropogenic acidification driven by N fertilization is indeed the main cause, being 10 to 100 times greater than that associated with acid deposition (Guo *et al.*, 2010).

### 2.2.3 | ACIDIFICATION AND CROP LOSS

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Acidification of agricultural soils may contribute to further soil pollution, through the mobilization of toxic heavy metals. If the content of nitrogen applied to agricultural soils is higher than the plants' requirements, nitrification microbial activity will lead to the accumulation of nitrates ( $\text{NO}_3^-$ ) that can easily leach to groundwater due to their high solubility, polluting it (Tian *et al.*, 2015). When soil nutrient availability increases, microbial biomass and activity increases as well, but the microbial biodiversity is altered, causing imbalances in the nutrient cycle (Lu and Tian, 2017).

The main risk from P fertilizers is transport to surface water bodies, which has been documented to cause eutrophication of aquatic ecosystems in many regions (Stork and Lyons, 2012; Syers, Johnson and Curtin, 2008). The P is transported to water bodies adsorbed to eroded soil particles or from excessive amounts of P fertilizer or animal manure applied when conditions are not suitable (Syers, Johnson and Curtin, 2008). Many farmlands receive more P inputs than the amount that crops are able to take up, causing a soil-P surplus, at least in the short term (Aarts, Habekotté and Keulen, 2000; Syers, Johnson and Curtin, 2008).

### 2.2.4 | PESTICIDES

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An extensive review of scientific research about the effects of pesticide use on soil functions was recently undertaken by the Intergovernmental Technical Panel on Soil (FAO and ITPS, 2017). The main scientific-based evidence presented in this work showed an increase in the farmers' net return when they applied pesticides, however the benefits of pesticide use are usually assessed by comparing use of synthetic pesticides versus no use of pesticides rather than comparing synthetic pesticides to biological control of pests (Cai, 2008). Negative associated impacts of specific pesticides on soil organisms and soil functions have been also reported.

For example, some organochlorine pesticides suppress symbiotic nitrogen fixation, resulting in lower crop yields (Fabra, 1997; Fox *et al.*, 2007; Santos and Flores, 1995). The FAO and ITPS report also highlights the knowledge gap on the relationship between pesticides and soil health, mainly on soil pollution (FAO and ITPS, 2017). The international efforts to assess the ecotoxicological risk of pesticides and to control their use and release in the environment, through the Rotterdam and Stockholm Conventions (UNEP, 1998, 2001), constitute an important achievement in preventing and controlling soil pollution, but more information is needed regarding their specific interactions with soil components, their mobility in the soil matrix and possible plant uptake, and their effect on crop production (Arias-Estévez *et al.*, 2008). Especially for low and middle income countries not every single compound within the great variability of pesticides available in the market has been analyzed for its ecotoxicological effects before authorization (Aktar, Sengupta and Chowdhury, 2009). For example pesticides that have been taken off the market in High Income Countries (HICs) due to their severe adverse effects on human health and the environment frequently remain registered in Low and Middle Income Countries (LMICs). As pesticides residues can be found throughout the entire ecosystem, pesticide monitoring programmes about the level of residues in soils, surface and groundwater as well as and drinking water but particularly in food items are very important. However, in many low and middle income countries monitoring programmes are inexistent due to their scarcity of regulation capacity (Brodesser *et al.*, 2006).

### 2.2.5 | MANURE

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Application of untreated manure may lead to heavy metal pollution, which not only results in adverse effects on various parameters relating to plant quality and yield, but also causes changes in the size, composition and activity of the microbial community (Yao, Xu and Huang, 2003) affecting nutrient cycling and reducing nutrient availability.

As previously discussed, a high proportion of antibiotics given to livestock is poorly assimilated in the animals' guts and is excreted in urine and faeces. Untreated manure can thus contain high amounts of veterinary antibiotics (VA) that can lead to a rapid increase in antibiotic resistance in soils (see Section 2.3.2). The fate and effects of antibiotics in soils have gained great attention in the last few years, motivated in part by the results of the O'Neill commission report (O'Neill, 2014), which estimates that antimicrobial resistant infections may become the leading cause of death in the world by 2050.

Most common intestinal pathogens that enter the soil with manure and faeces are *Salmonella*, *Campylobacter*, and *Escherichia coli* viruses. The pathogen levels decrease with time and with high temperatures that are reached during storage before land application (Garcia *et al.*, 2010). Once spread on the soil, pathogens can survive for several months or years.

## 2.2.6 | URBAN WASTES IN AGRICULTURE

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Considering that the positive effects of sewage sludge amendment – such as waste reduction, nutrient cycling, increase of soil fertility, improvement of soil structure and water holding capacity – are significantly more important than the negative effects, efforts should focus on reducing the content of pollutants in sewage sludge and wastewaters used for irrigation. As highlighted by Petrie *et al.*, the lack of knowledge on the fate of emerging pollutants and other pollutants present in wastewater and sewage sludge can be solved only by analyzing them before land application (Petrie, Barden and Kasprzyk-Hordern, 2015).

Composting and pretreatments reduce the content of contaminants and pathogen organisms present in urban waste before their application as amendments in soils, and provide an economical and environmentally friendly approach for stabilizing animal waste and converting it into a worthy organic fertilizer. Frequently, however, high levels of heavy metals such as Pb, Cd, Cu, Zn, Cr, Ni, and salts remain in the amendments and may affect soil properties and inhibit plant growth (Bolan *et al.*, 2014; Hargreaves, Adl and Warman, 2008; Stasinou and Zabetakis, 2013; Stratton, Barker and Rechcigl, 1995). The heterogeneous composition of biosolids produced in different wastewater treatment plants requires chemical and biological investigation prior to soil application or incorporation (Bauman-Kaszubaska and Sikorski, 2009; Bien, Neczaj and Milczarek, 2013). Limited bioavailability and crop uptake of metals from composted biosolids in comparison with untreated sewage sludge demonstrates the need for pretreatment before its application to soils (Smith, 2009).



## 2.3 | HUMAN HEALTH RISKS ASSOCIATED WITH SOIL POLLUTION

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Oliver and Gregory summarise six soil-related human health risks (Oliver and Gregory, 2015). Of these, three are related to soil pollution: risks from elemental contamination (e.g. As, Cd, Pb); organic chemical contamination (e.g. PCBs, PAHs, POPs); and pharmaceutical contamination (e.g. estrogen, antibiotics). The three other risks are from soil pathogens such as anthrax and prions, micronutrient deficiencies, and under-nutrition due to degraded soils.

Long-term impacts of soil pollution on human health and the environment are still unclear, and many efforts are underway to better understand the mechanisms involved in natural attenuation and the health impacts of toxic pollutants (Bernhardt and Gysi, 2016). Urban soils deserve special attention because anthropogenic activities are concentrated on those soils, and the exposure patterns are more complex due to interactions with other health determinants such as nutrition, air quality, and access to health services for illness prevention (WHO, 2013). However, non-urban areas are also subjected to many different sources of pollution, frequently from diffuse sources, which makes it difficult to trace and to estimate their extent and risk. Future efforts related to soil pollution control and remediation should include these areas in their risk assessment approaches.



### 2.3.1 | PATHWAYS OF EXPOSURE OF HUMANS TO SOIL POLLUTANTS AND THEIR EFFECTS ON HUMAN HEALTH

The main pollutants related to industrial, mining, urban and agricultural land uses have been widely discussed in the previous chapters. This section will focus on the soil pollutants that are the most relevant to human health and the risks associated with them.

The route of human exposure to a soil contaminant will vary depending on the contaminant itself and on the conditions and activities at a particular site (Shayler, McBride and Harrison, 2009). Generally, people can be exposed to contaminants present in soil through ingestion or through the consumption of plants or animals that have accumulated large amounts of soil pollutants (Khan *et al.*, 2015); through dermal exposure, from using spaces such as parks and gardens (Chaparro Leal, Guney and Zagury, 2018); or by inhaling soil contaminants that have been vapourized (Figure 15). Humans may also be affected as a result of secondary contamination of water supplies and from deposition of air contaminants (Science Communication Unit, University of the West of England., 2013); in some situations, soils play an important role as the source of contaminants in these two processes.

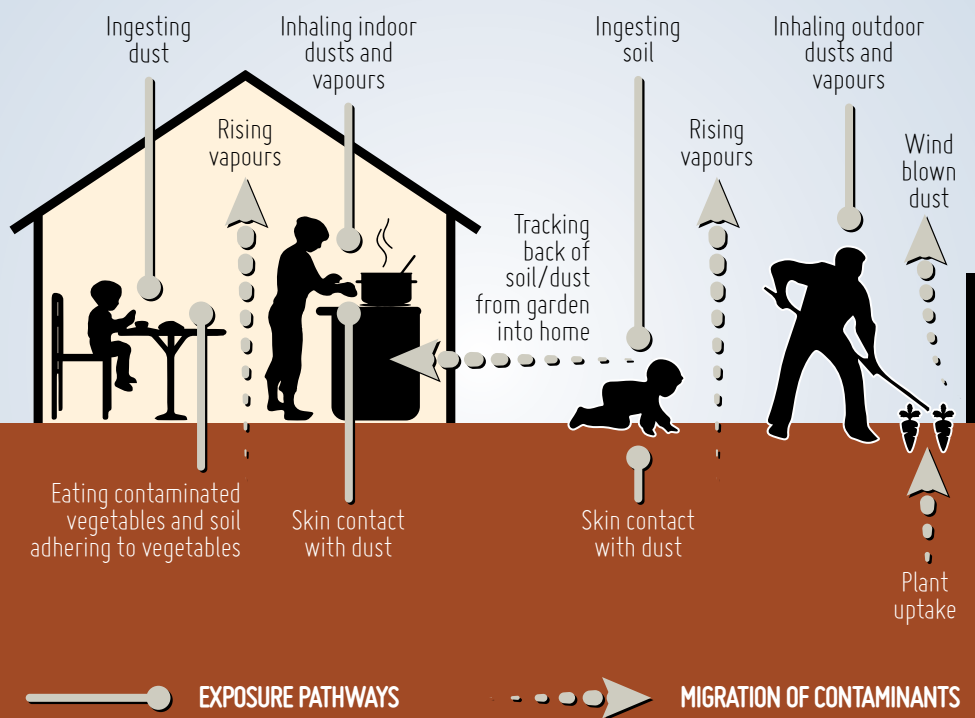


Figure 15. Possible exposure pathways of soil contamination in a residential scenario. Source: EA, 2008

Ten chemicals or groups of chemicals of major public health concern have been identified under the WHO's International Programme on Chemical Safety (WHO, 2010). The ten chemicals or groups of chemicals include soil pollutants such as, Cd, Pb and Hg; dioxin and dioxin-like substances and highly hazardous pesticides (HHP) - whose residues are transferred from contaminated soils to food and water bodies. HHPs are defined as those pesticides that are acknowledged to present particularly high levels of acute or chronic hazards to health or environment according to internationally accepted classification systems or their listing in relevant binding international agreements or conventions. In addition, pesticides that appear to cause severe or irreversible harm to health or the environment under conditions of use in a country may be considered to be and treated as highly hazardous (FAO and WHO, 2016). All of them have chronic effects due to long-term exposure and children, pregnant women and malnourished people are particularly vulnerable to pesticide exposure. Pathogens present in soil may also contaminate food, posing risks to human health. More than 200 diseases – ranging from diarrhea to cancers – are related to contaminated food intake (WHO, 2017b) and 24 percent of the world's population suffers from infections by soil-transmitted helminth, causing nutritional imbalance and chronic anemia (WHO, 2017a).

The uptake of metals by plants from soils may result in a significant risk to health (Brevik, 2013; Burgess, 2013; Jordão *et al.*, 2006). The absorption by plant roots is one of the main routes of entrance of heavy metals into the food chain, and varies according to the level of consumption (Pan *et al.*, 2010; Wagner, 1993). Cadmium and lead are the most toxic elements for man (Volpe *et al.*, 2009). Food is the main source of Cd intake by humans. A well-known case was in Japan, where ingestion of rice contaminated with Cd generated the disease known as itai-itai (Abrahams, 2002). Cadmium absorbed via food intake can penetrate through the placenta during pregnancy, damaging membranes and DNA and disrupting the endocrine systems, and can induce kidney, liver and bone damage (Brzóska and Moniuszko-Jakoniuk, 2005; Souza Arroyo *et al.*, 2012). The toxic effects of Pb affect several organs, causing biochemical imbalance in the liver, kidneys, spleen and lungs, and causing neurotoxicity, mainly in infants and children (Guerra *et al.*, 2012; Jaishankar *et al.*, 2014). Organomercuric compounds, especially methylmercury, are considered highly toxic. Mercury may induce changes in human neural and gastric systems and can lead to death. Arsenic is absorbed in the body orally or inhaled and is stored mainly in the liver, kidneys, heart and lungs, with smaller amounts accumulating in muscle and nerve tissue, and has been defined as carcinogenic (Brevik, 2013). It can lead to nervous systems disorders, liver and kidney failure as well as anemia and skin cancer. Nickel causes gastric, liver, and kidney defects and neurological effects (Brevik, 2013). Zinc is associated with anemia and tissue lesions, and while the negative effects of copper are rare, liver and kidney damage in infants is possible if exposure is prolonged (Brevik, 2013).

An increasing awareness in terms of the importance of vegetables and fruits to human diet and the identification of food as the main source for many contaminants suggest that the monitoring of heavy metals in food crops should be carried out frequently. The World Health Organization and FAO developed the Codex Alimentarius (WHO and FAO, 1995), which identifies safe limits for contaminants present in fruits, vegetables, fish and fishery products, and animal feed.

Aktar, Sengupta and Chowdhury presented a review of pesticide residues in food commodities in the European Union (Aktar, Sengupta and Chowdhury, 2009). Even though the amount of pesticides residues in food did not exceed the acceptable daily intake (ADI), few studies have analyzed the long term risk associated with these persistent pollutants in organisms (Kim, Kabir and Jahan, 2017; Xu *et al.*, 2017). Hernández *et al.* highlighted the need for further studies of pesticide mixture effects on human health, because current legislation considers maximum residue levels (MRL) of individual pesticides in food and water, without taking into account possible synergetic interactions at their low concentrations (Hernández *et al.*, 2013). Occupational exposure to pesticides is associated with various diseases including cancer, hormone disruption, asthma, allergies, and hypersensitivity (Burgess, 2013; Van Maele-Fabry *et al.*, 2010).

Intake of persistent organic pollutants accumulated in soils has a high relevance for human health (Figure 16). The last results of the WHO/UNEP global monitoring plan (GMP) show that levels of PCDDs, PCDFs and PCBs in human milk are still significantly above those considered toxicologically safe in many regions of the world, with a higher incidence in India and in some European and African countries (van den Berg *et al.*, 2017). Ingestion of soils (geophagia) has been a common practice in many African and South American countries (Woywodt and Kiss, 2002) among children and pregnant and breast-feeding women, and its practice has extended to western societies (Reeuwijk *et al.*, 2013). The intake of contaminated clay with POPs and heavy metals is a soil-borne source of diseases, as daily exposure levels are frequently exceeded (Odongo, Moturi and Mbutia, 2016). Bányiová *et al.* pointed out that the main source of POP exposure in the Czech Republic is through intake of polluted food (Bányiová *et al.*, 2017). Even though POP levels in human bodies have been reduced since the introduction of the Stockholm Convention, accidents still occur and they are an important source of soil and food contamination (Hilscherova *et al.*, 2007). The presence of POPs in human milk represents a high risk to the health of newborns and fetuses, as the POPs are circulating in the mother's body (Reeuwijk *et al.*, 2013).





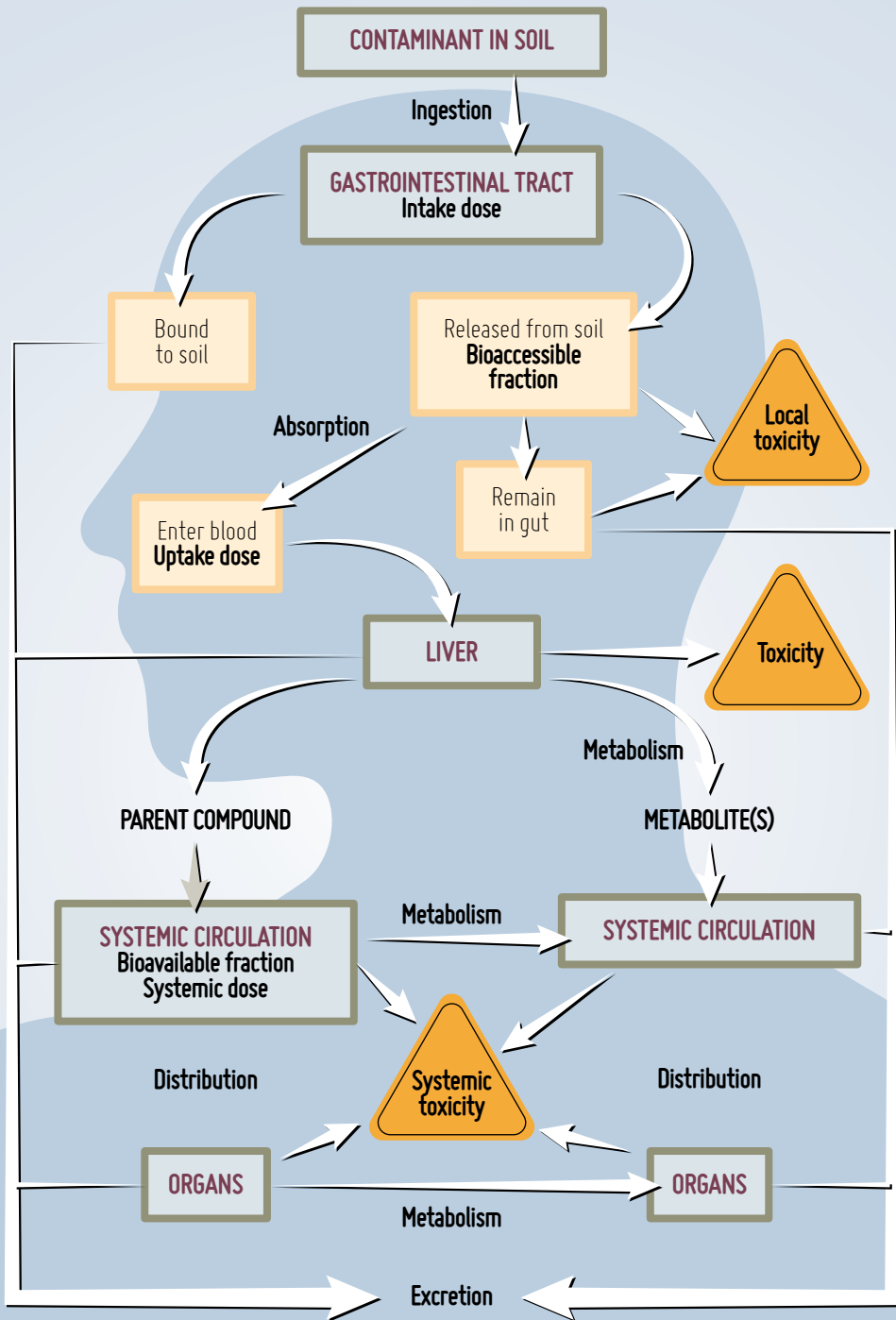


Figure 16. Simplified pathway for oral exposure to soil pollutants. Source: Hosford, 2008

The major exposure route for polycyclic PAHs is through contaminated food ingestion and they are a suspected carcinogenic risk (Brody *et al.*, 2007; Xia *et al.*, 2010). Due to the aromatic nature of PAHs, they easily penetrate cellular membranes and bind covalently with DNA molecules, where they may cause mutations (Muñoz and Albores, 2011). The establishment of a health risk assessment of PAHs would be complex because of the many uncertainties related to exposure and toxicity, which are still unsolved.

Regarding emerging pollutants, there are gaps in our understanding of how they behave in the environment, which interactions occur within the soil matrix, and their toxicity, bioaccumulation properties and transport mechanisms in human bodies, despite the already available information on exposure routes and levels in human tissues (Covaci *et al.*, 2011). Normally, these compounds appear at very low concentrations in human bodies, but the fact that many of them have only recently emerged as contaminants means that long-term studies that focus on the epidemiological aspects of this issue are needed.

Comprehensive assessments of the health effects of most forms of soil, heavy metals, and chemical pollution have not yet been published (Landrigan *et al.*, 2018). Basic toxicological data and knowledge on exposure routes and rates are needed in order to analyze the effects of soil pollutants on human health. A risk assessment based on a toxicological approach that considers tolerable doses is frequently used to establish hazard-related doses (Blume *et al.*, 2016; WHO, 2013). These acceptable doses are characterized for examining the harmful effects of individual substances on a person's health considering a conceptual exposure scenario that identifies sources, pathways and receptors.

As described in the WHO's report (WHO, 2013), there are several scientific tools that can be used to evaluate the risk posed to human health by soil pollution. Other models recently developed to assess the human exposure include BROWSE (Bystanders, Residents, Operators and WorkerS Exposure models for plant protection products) which contains more realistic scenarios (Butler Ellis *et al.*, 2017) and integrates large European guidance and regulatory databases to refine the assessment of human exposure (Lammoglia *et al.*, 2017). The real health risk assessment has not yet been well defined in the case of simultaneous exposure to two or more chemical substances, which occurs in real-life conditions and may have synergistic effects (Nicolopoulou-Stamati *et al.*, 2016).

### 2.3.2 | SOILS AS RESERVOIR OF ANTIMICROBIAL RESISTANT BACTERIA AND GENES

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The transference of antibiotic resistance genes from the environment to human pathogens has created a great challenge due to an overall decrease in effectivity of antibiotics (Harbarth *et al.*, 2015; Thomas and Nielsen, 2005; WHO, 2018). As a result, infections persist in the body, increasing the risk of contamination of others (WHO, 2018). Each year approximately 700 000 deaths occur globally that are attributable to AMR bacteria; 25 000 deaths in Europe (EC, 2017) and around 23 000 in the United States of America (CDC, 2013). Furthermore, human health implications for intake of antibiotic residues and AMR bacteria present in food are largely unknown, although several potential adverse impacts have been observed.

These include allergic and toxic reactions or chronic toxic effects as a result of prolonged low-level exposure (McManus *et al.*, 2002; Sarmah, Meyer and Boxall, 2006). The risks of AMR are especially important in newborns, where AMR bacteria populate the newborns' guts (Brinkac *et al.*, 2017).

Soil is considered to represent a natural reservoir of antibiotic-resistant bacteria carrying a diverse set of known and unknown resistance determinants (Cytryn, 2013). Fungi and bacteria that occur naturally in the environment produce many antibiotics that humans have been using for centuries, and at the same time they have antibiotic-resistant genes against the antibiotics they produce (Hopwood, 2007). Allochthonous resistant bacteria and genes added to soil with manure or sewage sludge might not be well adapted to soil conditions as they are subject to selective pressure by native organisms (Heuer *et al.*, 2008).

When microorganisms (such as bacteria, fungi, viruses and parasites) are continuously exposed to antibiotics or another antimicrobial agent that kills or inhibits the growth of microorganisms, selection of resistant organisms occurs, even at low concentrations. As well, transference of the resistome (extrachromosomal antibiotic-resistant plasmids) or of the mutated genes inside the chromosome to other members of their own species and to other species occurs (Khachatourians, 1998).

Antibiotics are used worldwide for animal therapy and growth promotion in livestock production. A high percentage of these antibiotics are not assimilated by organisms' systems and are excreted into the environment. Assimilation rates by livestock depend on the pharmacokinetics and transformation of the antibiotics by the animal's metabolism. Heuer *et al.*, for example, found that more than 96 percent of the veterinary antibiotic sulfadiazine was excreted in its parental form or as metabolites ten days after administration to pigs (Heuer *et al.*, 2008). Excretion rates are lower for tetracyclines (Winckler and Grafe, 2001), but can reach high rates, over 90 percent for amoxicillin and difloxacin (Sukul *et al.*, 2009). Large amounts of these antibiotics end up in farm and urban wastes and soils, and they are not completely removed in wastewater treatment plants or during composting processes. Wastewater treatment plants were identified as the main source of antibiotic release into streams, with highly variable removal rates for different antibiotics (Michael *et al.*, 2013; Watkinson *et al.*, 2009).

Changes in soil microorganism communities and in crucial activities have been observed after the application of amendments to soil, promoting resistant populations (Ding *et al.*, 2014; Tian *et al.*, 2015; Tien *et al.*, 2017). Antimicrobial-resistant populations in manure may be responsible for the horizontal transference of resistant plasmid to soil-dwelling organisms, as confirmed under field conditions in manure-amended soils (Gotz and Smalla, 1997; Smit *et al.*, 1991). This transference process is promoted by the increase of nutrient sources, which activate microbial activity and population density (Ding *et al.*, 2014). Scientific evidence suggests that the presence of heavy metals, mainly Cu and Zn, in soils contributes to the co-selection of AMR (Grass, Rensing and Solioz, 2011; Hölzel *et al.*, 2012; Wales and Davies, 2015; Yu *et al.*, 2017). Mutations induced in soil microorganisms may lead to a range of metabolic phenotypes, including variations in the ability to use different carbon, nitrogen or phosphate sources (Perkins and Nicholson, 2008), altering global geochemical cycles (Allen *et al.*, 2010).

Wales and Davis also found that antimicrobial selection was enhanced by heavy metals even in the absence of antimicrobial substances when they were exposed to other biocides, such as disinfectants and antiseptics (Wales and Davies, 2015). Similar results were found in residential soils of Australia, where the high content of heavy metals stimulated the proliferation of antimicrobial-resistant genes (Knapp *et al.*, 2017). Thus, AMR becomes an even greater challenge in polluted soils. More research is needed to assess the risk of AMR spreading via sewage sludge and manure amendments (Bondarczuk, Markowicz and Piotrowska-Seget, 2016).

Antibiotic concentrations decrease rapidly after their entry in the soil through different processes (Figure 17), such as plant uptake, leaching to groundwater, sequestration in organic complexes, sorption to clay minerals or biodegradation (Jechalke *et al.*, 2014; Kuppusamy *et al.*, 2018). Despite this, an increase in resistant bacteria was observed even when antibiotic concentrations were very low (Gullberg *et al.*, 2011). Sequestration and sorption of antibiotics in soils reduce their bioavailability but increase their permanence and persistence in the environment (Jechalke *et al.*, 2014). The uptake of antibiotics by plants has been widely reported (Boxall *et al.*, 2006), and has gained attention as the food chain may be directly contributing to the spread of antibiotic resistance (Du and Liu, 2012). Furthermore, antibiotics may cause inhibition of seed germination and reduce crop growth (Du and Liu, 2012). If antibiotic-rich soils suffer any alteration that leads to changes in organic matter concentration and conformation (Gulkowska *et al.*, 2013), the sequestered antibiotics may be released in their bioavailable forms (Rosendahl *et al.*, 2011).



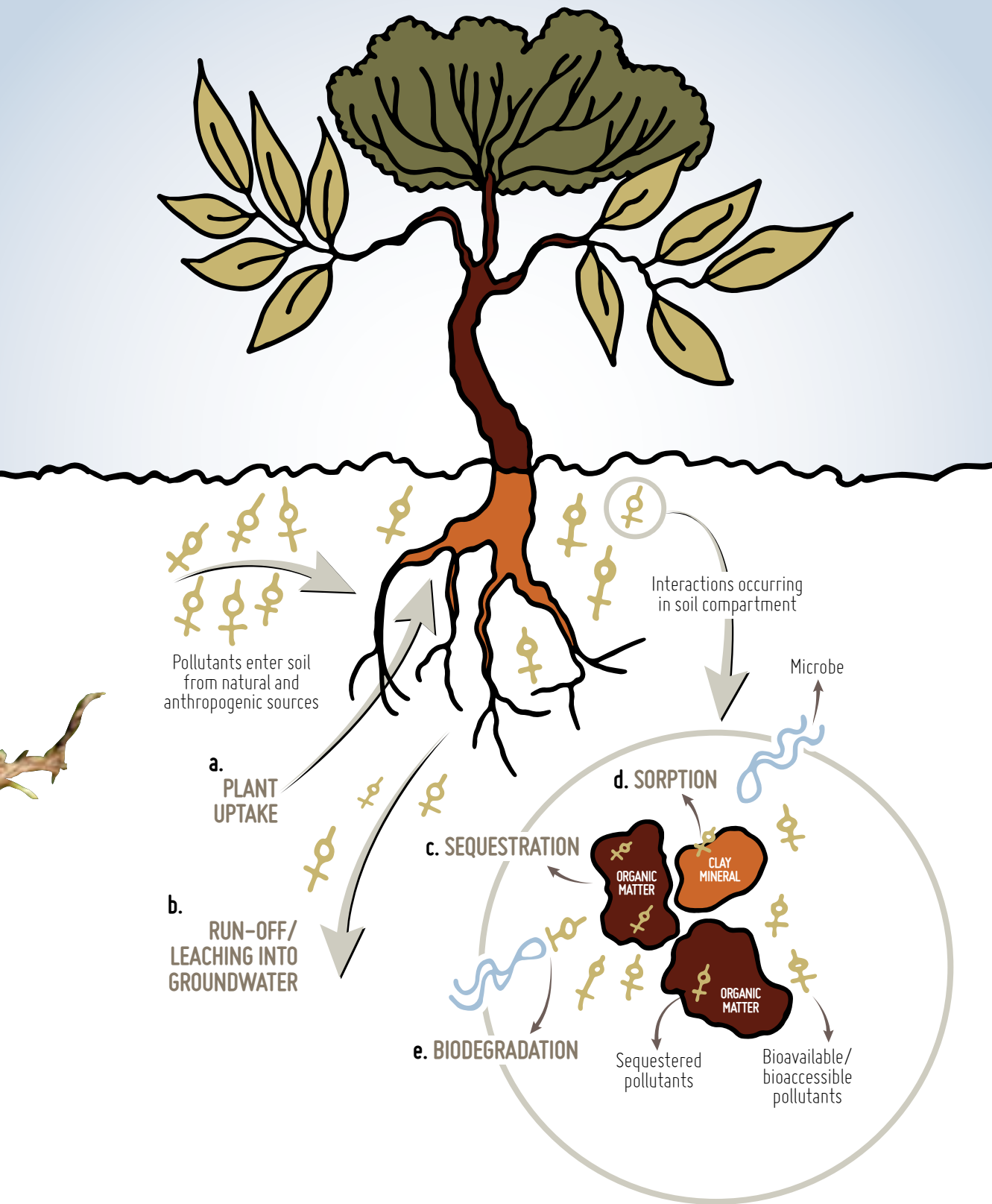


Figure 17. Fate of veterinary antibiotics in the environment. Source: Kuppusamy *et al.*, 2018

Different antibiotics have different target sites within cells; thus, organisms develop resistance to particular antibiotics and not a general resistance (Khachatourians, 1998). However, it has become more and more common to find multi-drug-resistant bacteria (CDC, 2013; EC, 2017; WHO, 2014). A recent study has shown a high co-presence of antibiotic-resistant genes and mobile genetic elements in a multi-drug-resistant bacteria community isolated from chicken manure (Yang *et al.*, 2017b). Bacteria present four main mechanisms that lead to antimicrobial resistance (Figure 18): 1) enzymatic degradation or modification of antibiotic compounds by intra and extracellular enzymes; 2) efflux pumps that actively remove antibiotic compounds outside the cell or in the periplasm; 3) modification or protection of the antibiotic binding site; and 4) natural or modified membrane permeability (Aleksun and Levy, 2007).

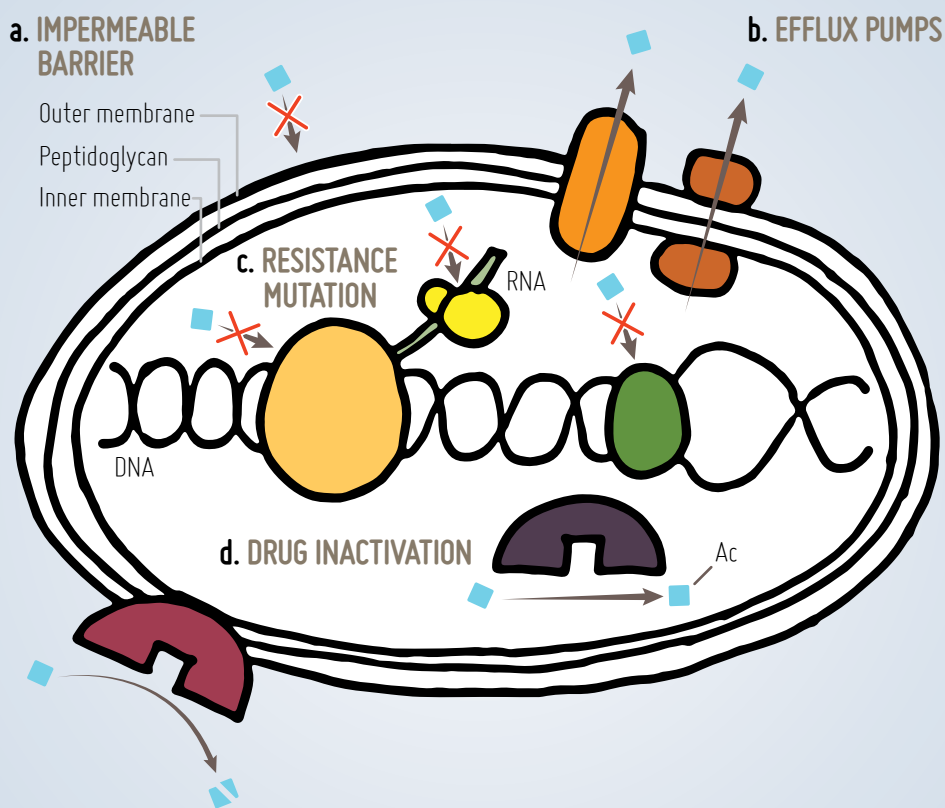


Figure 18. Main mechanisms of antimicrobial-resistance. Source: Allen *et al.*, 2010







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